

COMMON DOLPHIN (*Delphinus delphis delphis*): Western North Atlantic Stock

STOCK DEFINITION AND GEOGRAPHIC RANGE

The common dolphin (*Delphinus delphis delphis*) may be one of the most widely distributed species of cetaceans, as it is found world-wide in temperate and subtropical seas. In the North Atlantic, common dolphins are commonly found along the shoreline of Massachusetts in mass-stranding events (Bogomolni *et al.* 2010; Sharp *et al.* 2014). At-sea sightings have been concentrated over the continental shelf between the 100-m and 2000-m isobaths and over prominent underwater topography and east to the mid-Atlantic Ridge (29°W; Doksaeter *et al.* 2008; Waring *et al.* 2008). Common dolphins have been noted to be associated with Gulf Stream features (CETAP 1982; Selzer and Payne 1988; Waring *et al.* 1992; Hamazaki 2002). The species is less common south of Cape Hatteras, although schools have been reported as far south as the Georgia/South Carolina border (32° N; Jefferson *et al.* 2009). They exhibit seasonal movements, where they are found from Cape Hatteras northeast to Georges Bank (35° to 42°N) during mid-January to May (Hain *et al.* 1981; CETAP 1982; Payne *et al.* 1984), although some animals tagged and released after stranding in winters of 2010–2012 used habitat in the Gulf of Maine north to almost 44°N (Sharp *et al.* 2016). Common dolphins move onto Georges Bank, Gulf of Maine, and the Scotian Shelf from mid-summer to autumn. Selzer and Payne (1988) reported very large aggregations (greater than 3,000 animals) on Georges Bank in autumn. Migration onto the Scotian Shelf and continental shelf off Newfoundland occurs during summer and autumn when water temperatures exceed 11°C (Sergeant *et al.* 1970; Gowans and Whitehead 1995).

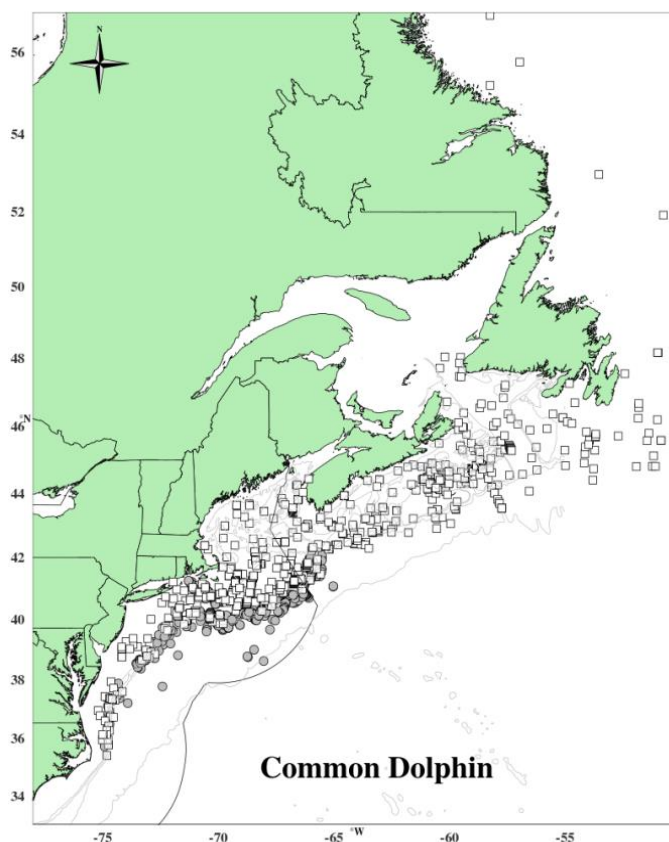


Figure 1. Distribution of common dolphin sightings from NEFSC and SEFSC shipboard and aerial surveys during the summers of 1998, 1999, 2002, 2004, 2006, 2007, 2010, 2011, 2016 and Department of Fisheries and Oceans Canada 2007 TNASS and 2016 NAISS surveys. Isobaths are the 100-m, 1000-m and 4000-m depth contours.

Westgate (2005) tested the proposed one-population-stock model using a molecular analysis of mitochondrial DNA (mtDNA), as well as a morphometric analysis of cranial specimens. Both genetic analysis and skull morphometrics failed to provide evidence ($p > 0.05$) of more than a single population in the western North Atlantic, supporting the proposed one-stock model. However, when western and eastern North Atlantic common dolphin mtDNA and skull morphology were compared, both the cranial and mtDNA results showed evidence of restricted gene flow ($p < 0.05$) indicating that these two areas are not panmictic. Cranial specimens from the two sides of the North Atlantic differed primarily in elements associated with the rostrum. These results suggest that common dolphins

in the western North Atlantic are composed of a single panmictic group whereas gene flow between the western and eastern North Atlantic is limited (Westgate 2005, 2007). This was further supported by Mirimin *et al.* (2009) who investigated genetic variability using both nuclear and mitochondrial genetic markers and observed no significant genetic differentiation between samples from within the western North Atlantic region, which may be explained by seasonal shifts in distribution between northern latitudes (summer months) and southern latitudes (winter months). However, the authors point out that some uncertainty remains if the same population was sampled in the two different seasons.

POPULATION SIZE

The current best abundance estimate for Western North Atlantic stock of common dolphins is 172,947 (CV=0.21) which is the total of Canadian and U.S. surveys conducted in 2016 (Table 1). This estimate, derived from shipboard and aerial surveys, covers most of this stock's known range. Because the survey areas did not overlap, the estimates from the three surveys were added together and the CVs pooled using a delta method to produce a species abundance estimate for the stock area.

Earlier Abundance Estimates

Please see Appendix IV for a summary of abundance estimates, including earlier estimates and survey descriptions. As recommended in the guidelines for preparing Stock Assessment Reports (NMFS 2016), estimates older than eight years are deemed unreliable to determine a current PBR.

Recent Surveys and Abundance Estimates

Abundance estimates of 48,723 (CV=0.48) for the Newfoundland/Labrador portion and 43,124 (CV=0.28) for the Bay of Fundy/Scotian Shelf/Gulf of St. Lawrence portion of the stock area were generated from the Canadian Northwest Atlantic International Sightings Survey (NAISS) survey conducted in August–September 2016 (Table 1). This large-scale aerial survey covered Atlantic Canadian shelf and shelf break habitats from the northern tip of Labrador to the U.S. border off southern Nova Scotia (Lawson and Gosselin 2018). Line-transect density and abundance analyses were completed using Distance 7.1 release 1 (Thomas *et al.* 2010).

Abundance estimates of 80,227 (CV=0.31) and 900 (CV=0.57) common dolphins were generated from vessel surveys conducted in U.S. waters of the western North Atlantic during the summer of 2016 (Table 1; Garrison 2020; Palka 2020). One survey was conducted from 27 June to 25 August in waters north of 38°N latitude and consisted of 5,354 km of on-effort trackline along the shelf break and offshore to the outer limit of the U.S. EEZ (NEFSC and SEFSC 2018). The second vessel survey covered waters from Central Florida to approximately 38°N latitude between the 100-m isobaths and the outer limit of the U.S. EEZ during 30 June–19 August. A total of 4,399 km of trackline was covered on effort (NEFSC and SEFSC 2018). Both surveys utilized two visual teams and an independent observer approach to estimate detection probability on the trackline (Laake and Borchers 2004). Mark-recapture distance sampling was used to estimate abundance. Estimates from the two surveys were combined and CVs pooled to produce a species abundance estimate for the stock area.

Table 1. Summary of recent abundance estimates for western North Atlantic common dolphin (*Delphinus delphis delphis*) by month, year, and area covered during each abundance survey, and resulting abundance estimate (*N_{est}*) and coefficient of variation (CV). The estimate considered best is in bold font.

Month/Year	Area	Nest	CV
June–Sep 2016	Central Virginia to lower Bay of Fundy	80,227	0.31
June–Aug 2016	Florida to Central Virginia	900	0.57
June–Sep 2016	Newfoundland/Labrador	48,723	0.48
June–Sep 2016	Bay of Fundy/Scotian Shelf/Gulf of St. Lawrence	43,124	0.28
June–Sep 2016	Florida to Newfoundland/Labrador (COMBINED)	172,974	0.21

Minimum Population Estimate

The minimum population estimate is the lower limit of the two-tailed 60% confidence interval of the log-normally distributed best abundance estimate. This is equivalent to the 20th percentile of the log-normal distribution as specified by Wade and Angliss (1997). The best estimate of abundance for common dolphins is 172,974 animals (CV=0.21), derived from the 2016 aerial and shipboard surveys. The minimum population estimate for the western North Atlantic

common dolphin is 145,216.

Current Population Trend

A trend analysis has not been conducted for this stock. The statistical power to detect a trend in abundance for this stock is poor due to the relatively imprecise abundance estimates and long survey interval (see Appendix IV for a survey history of this stock). For example, the power to detect a precipitous decline in abundance (i.e., 50% decrease in 15 years) with estimates of low precision (e.g., $CV > 0.30$) remains below 80% ($\alpha = 0.30$) unless surveys are conducted on an annual basis (Taylor *et al.* 2007). There is current work to standardize the strata-specific previous abundance estimates to consistently represent the same regions and include appropriate corrections for perception and availability bias. These standardized abundance estimates will be used in state-space trend models that incorporate environmental factors that could potentially influence the process and observational errors for each stratum.

CURRENT AND MAXIMUM NET PRODUCTIVITY RATES

Due to uncertainties about the stock-specific life-history parameters, the maximum net productivity rate was assumed to be the default value for cetaceans of 0.04. This value is based on theoretical modeling showing that cetacean populations may not grow at rates much greater than 4% given the constraints of their reproductive life history (Barlow *et al.* 1995).

POTENTIAL BIOLOGICAL REMOVAL

Potential Biological Removal (PBR) is the product of minimum population size, one-half the maximum productivity rate, and a recovery factor (MMPA Sec. 3. 16 U.S.C. 1362; Wade and Angliss 1997). The minimum population size is 145,216 animals. The maximum productivity rate is 0.04, the default value for cetaceans. The recovery factor is 0.5, the default value for stocks of unknown status and with the CV of the average mortality estimate less than 0.3 (Wade and Angliss 1997). PBR for the western North Atlantic stock of common dolphin is 1,452.

Table 2. Best and minimum abundance estimates for the western North Atlantic common dolphin (*Delphinus delphis delphis*) with Maximum Productivity Rate (R_{max}), Recovery Factor (Fr) and PBR.

Nest	CV	Nmin	Fr	Rmax	PBR
172,974	0.21	145,216	0.5	0.04	1,452

ANNUAL HUMAN-CAUSED MORTALITY AND SERIOUS INJURY

Average annual estimated fishery-related mortality or serious injury to this stock during this reporting period are presented in Table 3.

Table 3. The total annual estimated average human-caused mortality and serious injury for the western North Atlantic common dolphin (*Delphinus delphis delphis*).

Years	Source	Annual Avg.	CV
2015–2019	U.S. fisheries using observer data	390	0.11
2015–2019	Research mortalities	0.2	
2015–2019	Non-fishery stranding mortalities	0.2	
TOTAL		390.4	

Uncertainties not accounted for include the potential that the observer coverage was not representative of the fishery during all times and places. There are no major known sources of unquantifiable human-caused mortality or serious injury for this stock.

Northeast Sink Gillnet

Annual common dolphin mortalities were estimated using annual ratio-estimator methods (Orphanides and Hatch 2017; Orphanides 2019, 2020, 2021; Precoda and Orphanides 2022). See Table 4 for bycatch estimates and observed mortality and serious injury for the current 5-year period, and Appendix V for historical bycatch information.

Mid-Atlantic Gillnet

Common dolphins were taken in observed trips during most years. Annual common dolphin mortalities were estimated using annual ratio-estimator methods (Orphanides and Hatch 2017; Orphanides 2019, 2020, 2021; Precoda and Orphanides 2022). See Table 4 for bycatch estimates and observed mortality and serious injury for the current 5-year period, and Appendix V for historical bycatch information.

Northeast Bottom Trawl

This fishery is active in New England waters in all seasons. Annual common dolphin mortalities were estimated using annual stratified ratio-estimator methods (Lyssikatos and Chavez-Rosales 2022). See Table 4 for bycatch estimates and observed mortality and serious injury for the current 5-year period, and Appendix V for historical bycatch information.

Mid-Atlantic Bottom Trawl

Annual common dolphin mortalities were estimated using annual stratified ratio-estimator methods (Lyssikatos and Chavez-Rosales 2022). See Table 4 for bycatch estimates and observed mortality and serious injury for the current 5-year period, and Appendix V for historical bycatch information.

Pelagic Longline

Pelagic longline bycatch estimates of common dolphins for 2015–2019 were documented in Garrison and Stokes (2017, 2020a, 2020b, 2021). There is a high likelihood that dolphins released alive with ingested gear or gear wrapped around appendages will not survive (Wells *et al.* 2008). See Table 4 for bycatch estimates and observed mortality and serious injury for the current 5-year period, and Appendix V for historical bycatch information.

Research Takes

In October 2016 the University of Rhode Island, Graduate School of Oceanography reported the incidental capture/drowning of a 206-cm female common dolphin during a routine, weekly research trawl fishing trip in Narragansett Bay, Rhode Island. The incident was reported to Mystic Aquarium, Mystic, Connecticut; NOAA GARFO Office, Gloucester, Massachusetts; NOAA law enforcement; and NOAA Protected Species Branch, Woods Hole, Massachusetts. A complete necropsy was conducted at the Wood Hole Oceanographic Institution, Woods Hole, Massachusetts.

Table 4. Summary of the incidental serious injury and mortality of North Atlantic common dolphins (*Delphinus delphis delphis*) by commercial fishery including the years sampled, the type of data used, the annual observer coverage, the serious injuries and mortalities recorded by on-board observers, the estimated annual serious injury and mortality, the combined serious injury and mortality estimate, the estimated CV of the annual combined serious injury and mortality and the mean annual serious injury and mortality estimate (CV in parentheses).

Fishery	Years	Data Type ^a	Observer Coverage ^b	Observed Serious Injury ^d	Observed Mortality	Estimated Serious Injury ^d	Estimated Mortality	Estimated Combined Mortality	Estimated CVs	Mean Combined Annual Mortality
Northeast Sink Gillnet	2015	Obs. Data,	0.14	0	3	0	55	55	0.54	73 (0.19)
	2016	Trip	0.10	0	8	0	80	80	0.38	
	2017	Logbook,	0.12	0	20	0	133	133	0.28	
	2018	Allocated	0.11	0	10	0	93	93	0.45	
	2019	Dealer Data	0.13	0	1	0	5.0	5.0	0.68	
Mid-Atlantic Gillnet	2015	Obs.	0.06	0	3	0	30	30	0.55	17 (0.31)
	2016	Data,	0.08	0	1	0	7	7	0.97	
	2017	Trip	0.09	1	1	11	11	22	0.71	
	2018	Weighout	0.09	0	1	1	7.7	7.7	0.91	
	2019		0.13	0	3	0	20	20	0.56	
Northeast Bottom Trawl ^c	2015	Obs.	0.19	0	4	0	22	22	0.45	15 (0.27)
	2016	Data,	0.12	0	2	0	16	16	0.46	
	2017	Trip	0.16	0	0	0	0	0	0	
	2018	Logbook	0.12	0	4	0	28	28	0.54	
	2019		0.16	0	2	0	10	10	0.62	
Mid-Atlantic	2015	Obs.	0.09	0	26	0	250	250	0.32	281 (0.12)
	2016	Data,	0.10	0	22	0	177	177	0.33	

Bottom Trawl ^c	2017	Dealer Data	0.10	0	66	0	380	380	0.23	
	2018		0.12	1	34	5	200	205	0.54	
	2019		0.12	2	52	15	395	395	0.23	
Pelagic Longline	2015	Obs. Data,	0.12	1	0	9.05	0	9.05	1	3.1 (0.67)
	2016	Logbook Data	0.15	0	0	0	0	0	0	
	2017		0.12	1	0	4.92	0	4.92	1	
	2018		0.10	1	0	1.44	0	1.44	1	
	2019		0.10	0	0	0	0	0	0	
TOTAL										390 (0.11)

a. Observer data (Obs. Data), used to measure bycatch rates, are collected within the Northeast Fisheries Observer Program and At-sea Monitoring Program. NEFSC collects landings data (unallocated Dealer Data or Allocated Dealer Data) which are used as a measure of total landings and mandatory Vessel Trip Reports (VTR; Trip Logbook) are used to determine the spatial distribution of landings and fishing effort.

b. Observer coverage is defined as the ratio of observed to total metric tons of fish landed for the gillnet fisheries and the ratio of observed to total trips for bottom trawl and Mid-Atlantic mid-water trawl (including pair trawl) fisheries.

c. Fishery related bycatch rates were estimated using an annual stratified ratio-estimator (Lyssikatos and Chavez-Rosales 2022).

d. Serious injuries were evaluated for the period and include both at-sea monitor and traditional observer data (Josephson *et al.* 2022)

Other Mortality

Common dolphins reported stranded between Maine and Florida are reported in Table 5 (NOAA National Marine Mammal Health and Stranding Response Database unpublished data, accessed 17 November 2020). The total includes mass-stranded common dolphins in Massachusetts during 2015 (a total of 37 in 13 events), 2016 (a total of 35 animals in 9 events), 2017 (over 90 animals in 20 events), and 2018 (a total of 28 animals in 9 events) and 2019 (28 animals in 9 events). Animals released or last sighted alive include 9 in 2015, 23 in 2016, 70 in 2017, 18 in 2018 and 4 in 2019. In 2015, 2 cases were classified as human interactions, both in Rhode Island, and both related to mutilation likely to be post-mortem. Seven cases in 2016 were coded as human interaction. All but 2 of these were released alive. One of the 2 was a fishery interaction and the other was coded HI (Human Interaction) due to a beachgoer intervention. Six cases in 2017 were coded as human interaction, 2 of which were classified as fishery interactions, 1 classified as a possible boat collision, and 1 released alive. Another dolphin was euthanized after multiple restrandings and another was HI due to beachgoer intervention. In 2018, 5 cases were coded as human interactions. Two were public harassment and 3 involved fishing gear, though only one was classified as a fishery interaction. Eight stranding mortalities in Massachusetts in 2019 were classified as human interactions and one each in New York and Rhode Island. The New York case was a fishery interaction. All were either coded as unlikely or undetermined that the HI contributed to the stranding. In this 5-year period, only 1 interaction (boat strike in 2017) was likely a non-fishery human-caused mortality. In an analysis of mortality causes of stranded marine mammals on Cape Cod and southeastern Massachusetts between 2000 and 2006, Bogomolni (2010) reported that 61% of stranded common dolphins were involved in mass-stranding events, and 37% of all the common dolphin stranding mortalities were disease-related.

The Marine Animal Response Society of Nova Scotia reported 2 common dolphins stranded in 2015, 5 in 2016, 5 in 2017, 5 in 2018, and 4 in 2019 (Tonya Wimmer/Andrew Reid, pers. comm.).

Table 5. Common dolphin (*Delphinus delphis delphis*) reported strandings along the U.S. Atlantic coast, 2015–2019.

STATE	2015	2016	2017	2018	2019	TOTALS
New Hampshire	1	1	2	0	0	4
Massachusetts ^a	40	67	166	61	95	429
Rhode Island ^b	7	4	5	4	5	25
Connecticut	2	1	1	0	0	4
New York	3	3	15	11	9	41
New Jersey	3	5	0	2	4	14
Delaware	2	0	0	0	1	3

Maryland	1	0	0	0	2	3
Virginia	2	0	1	3	5	11
North Carolina	4	1	0	3	4	12
TOTALS	65	82	190	84	125	546

It should be recognized that evidence of human interaction does not always indicate cause of death, but rather only that there was evidence of interaction with a fishery (e.g., line marks, net marks) or evidence of a boat strike, gunshot wound, mutilation, etc., at some point, including post-stranding. Stranding data probably underestimate the extent of mortality and serious injury because all of the marine mammals that die or are seriously injured may not wash ashore, nor will all of those that do wash ashore necessarily show signs of entanglement or other fishery interaction. Finally, the level of technical expertise among stranding network personnel varies widely as does the ability to recognize signs of fishery interaction. However a recently published human interaction manual (Barco and Moore 2013) and case criteria for human interaction determinations (Moore *et al.* 2013) should help with this.

HABITAT ISSUES

The chronic impacts of contaminants (polychlorinated biphenyls [PCBs] and chlorinated pesticides [DDT, DDE, dieldrin, etc.]) on marine mammal reproduction and health are of concern (e.g., Pierce *et al.* 2008; Jepson *et al.* 2016; Hall *et al.* 2018; Murphy *et al.* 2018), but research on contaminant levels for the western north Atlantic stock of common dolphins is lacking.

Anthropogenic sound in the world's oceans has been shown to affect marine mammals, with vessel traffic, seismic surveys, and active naval sonars being the main anthropogenic contributors to low- and mid-frequency noise in oceanic waters (e.g., Nowacek *et al.* 2015; Gomez *et al.* 2016; NMFS 2018). The long-term and population consequences of these impacts are less well-documented and likely vary by species and other factors. Impacts on marine mammal prey from sound are also possible (Carroll *et al.* 2017), but the duration and severity of any such prey effects on marine mammals are unknown.

Climate-related changes in spatial distribution and abundance, including poleward and depth shifts, have been documented in or predicted for plankton species and commercially important fish stocks (Nye *et al.* 2009; Head *et al.* 2010; Pinsky *et al.* 2013; Poloczanska *et al.* 2013; Hare *et al.* 2016; Grieve *et al.* 2017; Morley *et al.* 2018) and cetacean species (e.g., MacLeod 2009; Sousa *et al.* 2019). There is uncertainty in how, if at all, the distribution and population size of this species will respond to these changes and how the ecological shifts will affect human impacts to the species.

STATUS OF STOCK

Common dolphins are not listed as threatened or endangered under the Endangered Species Act, and the Western North Atlantic stock is not considered strategic under the Marine Mammal Protection Act. The 2015–2019 average annual human-related mortality does not exceed PBR. The total U.S. fishery-related mortality and serious injury for this stock is not less than 10% of the calculated PBR and, therefore, cannot be considered to be insignificant and approaching zero mortality and serious injury rate. The status of common dolphins, relative to Optimum Sustainable Population (OSP), in the U.S. Atlantic EEZ is unknown.

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