# Appendix A Southern DPS Green Sturgeon Recovery Plan

Threats Assessment Methodology and References

# **Background and Definitions**

This Appendix describes the threats assessment process and methodology. The threats assessment is to determine, to the extent possible, why the species is declining. For the purposes of this recovery plan, a threat is defined as any factor that could represent an impediment to recovery. Understanding of current and potential future threats to sDPS green sturgeon is essential in developing effective recovery actions and criteria.

The threats assessment was separated into habitat units and conducted by recovery team members who had experience with a particular habitat unit as follows:

**Sacramento River Basin (SRB):** 1) upstream extent of the Sacramento-San Joaquin Delta on the Sacramento River (I Street Bridge; defined by California Water Code section 12220) to Keswick Dam, including the Sutter and Yolo bypasses; 2) the Feather River from its confluence with the Sacramento River upstream to Fish Barrier Dam; 3) the Yuba River from its confluence with the Feather River upstream to Daguerre Point Dam; and 4) the American River from its confluence with the Sacramento River upstream to the Highway 160 bridge (Recovery Team members: Corwin, Poytress, Seesholtz).

**San Francisco Bay Delta and Estuary (SFBDE):** estuarine areas up to the top of high tide from the Golden Gate Bridge to the upstream extent of the Sacramento-San Joaquin Delta (excluding flood control bypasses) (Recovery Team members: Gingras, Israel).

**Coastal Bays and Estuaries (CBE):** coastal bays and estuaries up to top of high tide including: 1) Humboldt Bay in California; 2) Coos, Winchester, Yaquina, and Nehalem bays in Oregon; 3) Willapa Bay and Grays Harbor in Washington; and 4) the lower Columbia River estuary from the mouth to river kilometer 74 (Recovery Team members: Erickson, Moser, Parsley).

**Nearshore Marine (NM):** nearshore waters within the 60 fathom (110 meters) isobath from the Monterey Bay north to the U.S./Canada border (including the Strait of Juan de Fuca) (Recovery Team members: Erickson, Moser, Parsley).

With the exception of the CBE and NM habitat units, each threats assessment was conducted independently.

The assessments were also conducted for each life stage present in a habitat unit (Table A-1).

Table A-1. Target life stage by habitat unit. Life stages were defined as follows: Eggs: fertilization to hatch; Larvae: hatch to size at metamorphosis (1 to 6 centimeters [cm] total length [TL]); Juveniles: metamorphosed juveniles to size at first ocean entry (6 to 65 cm TL); Subadults: first ocean entry to size at sexual maturity (65 to 150 cm TL); Adults: sexually mature adults (greater than 150 cm TL).

Habitat Unit	Life Stage					
	Adult	Subadult	Juveniles	Larvae	Eggs	
SRB	Х		Х	X	Х	
SFBDE	Х	Х	Х			
CBE	Х	Х				
NM	Х	Х				

Specific threats were organized with respect to each of the four geographic habitat units by Listing Factor and threat category as follows:

- A. Habitat Destruction, Modification, or Curtailment: threat categories Altered Water Flow, Altered Prey Base, Altered Water Temperature, Contaminants, Altered Sediment, Barriers to Migration, Water Depth Modification, Loss of Wetland Function, and Altered Turbidity.
- B. Overutilization for Recreational, Commercial, Scientific, or Educational Purposes: threat categories Take (harass, harm, pursue, hunt, shoot, wound, kill, trap, capture, or collect, or to attempt to engage in any such conduct) and Reduced Genetic Diversity.
- C. Disease and Predation: threat categories Disease and Predation.
- D. Inadequacy of Existing Regulatory Mechanisms: addressed in other listing factors and threat categories<sup>1</sup>.
- E. Other Natural or Man-made Factors: Competition for Habitat and Take.

The recovery team met and developed a list of current and potential future threats to the sDPS of green sturgeon. The team members determined the life stages, threat categories, and specific threats that would be assessed in their respective habitat unit. The following threat definitions were used:

**Anthropogenic light:** sources of anthropogenic light include but are not limited to construction sites (urban or industrial), bridges, boats, buoys, marinas, docks, and dams. Anthropogenic light may increase predation on all sizes of green sturgeon or alter behavior and migration.

Anthropogenic underwater sound: sources of underwater sound include but are not limited to cars or trains travelling over bridges, pile driving, blasting, boat engines, water pumps, air guns and hydrokinetic equipment. Effects on green sturgeon range from delayed and/or altered migration to stress, injuries, or death depending on the size of the fish and the amount of explosive, size of pile/hammer, or size/number of air guns.

<sup>&</sup>lt;sup>1</sup> Regulatory mechanisms were considered when ranking the threats under Listing Factors A through C and E.

**Aquaculture:** cultivation of aquatic organisms such as fish, crustaceans, and mollusks. Toxic contaminants may be linked to aquaculture activities occurring in bays, estuaries, and the nearshore marine environment. For example, pesticides are directly applied within intertidal areas of coastal bays to control burrowing shrimp or non-native grasses in an effort to promote shellfish culture.

**Artificial propagation of green sturgeon:** artificial propagation and release of green sturgeon for enhancement or research purposes. Addition of hatchery fish to the wild green sturgeon population may affect natural genotype frequencies and later population equilibrium. This could result in reduced levels of genetic diversity compared to the natural equilibrium. The overrepresentation of a particular genotype may lead to reduced resilience in the face of bottlenecks should population numbers continue to decline.

**Augmentation:** active and passive addition of gravel and other materials to enhance salmonid spawning habitat or promote shellfish habitat or culture. Gravel augmentation has the potential to affect depth structure and bottom composition, potentially affecting green sturgeon spawning habitat. Gravel and oyster shell placement occurs in some estuaries for sediment stabilization and to promote shellfish culture, thus altering substrates.

**Beach renourishment:** relocation of sands (usually as beneficial re-use of dredged material) on beaches for renourishment. Beach renourishment activities may alter sediment composition and transport in estuaries, reduce wetland function, and temporarily increase turbidity in nearshore marine waters.

**Bottom trawling:** trawling in groundfish and non-groundfish fisheries (e.g., shrimp, California halibut, ridgeback prawn, sea cucumber). Bottom trawling may affect habitat and cause prey to relocate. Discard of dead bycatch may affect the prey base and remove competing predators (potential prey base enhancements).

**Bypasses:** large flood control areas adjacent to the Sacramento River. The two major bypasses are Sutter and Yolo. Bypasses operate when flood waters in the Sacramento River overtop weirs located along the Sacramento River upstream of the Feather River and alter flood flows in the mainstem Sacramento River. Because bypasses are large shallow areas that can allow greater equilibration of water and air temperatures than adjacent river channels, the return water from the bypasses may be warmer or cooler than the receiving river water. Bypass weirs and other associated structures may delay or prevent green sturgeon migration. Adult green sturgeon are also stranded and exposed to dewatered areas and vulnerable to poaching around bypass weirs.

**Channel control structures:** structures used to control channel position including but not limited to levees, rip-rap, hard points, boulder weirs (e.g., Sunset Pumps), gradient control facilities (e.g., Glenn Colusa Irrigation District structures), wing dams, and pile dikes. Their presence alters water flow velocity and direction, channel migration and morphology, and sediment transport. These structures may also cause aggradation and degradation of channels (alteration of channel depths) used by green sturgeon for holding and spawning.

**Derelict fishing gear:** abandoned fishing gear (e.g., derelict gill nets in Puget Sound). Green sturgeon may be injured or killed through exposure to derelict fishing gear.

**Diversions:** major diversions include the south Delta pumping plants, Sacramento River facilities (Red Bluff Diversion Pumping Plant, Anderson Cottonwood Irrigation District, and Glenn Colusa Irrigation District), as well as facilities associated with Sunset Pumps and Daguerre Point Dam on the Feather and Yuba rivers respectively. Many unscreened smaller diversions are also operated throughout the system. Diversions reduce flow volume below facilities and potentially affect the magnitude of river hydraulics and stage in green sturgeon migration and spawning habitat, including flow into the San Francisco Bay-Delta/Estuary. This may affect migration (e.g., altering hydrology cues), water temperatures, and sediments. For example, reduced downstream flows may result in increased downstream water temperature and affect channel morphology and the potential scouring process of deep pool formation. Increased flows upstream of the diversion may result in decreased upstream water temperature (Also see "Upstream diversions").

Dredging and disposal of dredged material: sediment removal and/or disposal associated with maintaining navigable waters (e.g., marinas and boat ramps), construction projects (e.g., building bridges), and gold mining. Dredging can remove substrates important for spawning and rearing, alter benthic communities by removing benthic prey organisms or disrupting habitat, alter water depth, increase contaminant availability through exposure of sediment layers with elevated contaminant concentrations and dispersal of dredged material into the water column, and potentially result in suspension and release of fine-grain sediments (3-5% of total volume dredged by clamshell equipment) into the water column. The magnitude of this effect depends on sediment type and dredging objectives. Dredging has also been found to reduce the ability of sediments to remove phosphorous, nitrogen, and pesticides from the water column and releases phosphorous, nitrogen, and pesticides to clean water more quickly than sediments taken prior to dredging. Redistribution of mercury or mercury byproducts or other toxic legacy elements (e.g., arsenic, benzene, copper, and zinc) can occur from excavation activities. California has recently proposed to not permit suction gold mining dredging in the mainstem Sacramento River or its tributaries. Recovery of an area from dredging has been found to be variable dependent on sediment type. In-water disposal of dredged material can bury benthic communities, alters depth and bottom characteristics (further altering benthic communities), and disperses potentially contaminated sediments or fine grain material into the water column. For example, a disposal site near Alcatraz Island in central San Francisco Bay is no longer a deep depression. Dredged material was historically used to create levees in the Delta and thus contributed to the loss of wetlands. Currently, disposal of some dredged material is used beneficially and to a large extent for wetland creation (i.e., salt pond restoration). Temporary increases in turbidity could occur and depends upon the amount and type of material being disposed per event, the time between events, and the conditions (e.g., current) at the disposal site.

**Electromagnetic field:** electromagnetic fields generated by structures or facilities including but not limited to buried electrical cables, overhead power lines, and hydrokinetic projects. The presence of electromagnetic fields may delay or alter green sturgeon migration or affect their ability to forage, hold in pools, and/or spawn.

**Entrainment and/or impingement at water diversion intakes:** movement or transport of green sturgeon along with the flow of water (entrainment) and/or significant contact with diversion apparatus (impingement) at power plants and water diversion facilities. Larval and juvenile green sturgeon may suffer injury or mortality when they are entrained in unscreened or inadequately screened diversions and impinged on screened diversions. Attraction flows from large diversion facilities may alter migration or attract green sturgeon into suboptimal habitats (e.g., areas with increased predation or poor water quality). Mortality also may occur during collection, handling, and transport of salvaged green sturgeon (e.g., CVP and SWP pumping facilities). In marine waters, green sturgeon may be entrained at water intakes (e.g., at the Moss Landing power plant intake).

**Entrainment from dredging:** entrainment of green sturgeon associated with hydraulic dredging. White sturgeon have been entrained during hydraulic dredging in the Columbia River. Recent studies have shown that white sturgeon can be entrained during hydraulic dredging in the Delta.

**Entrainment from hydrokinetic projects:** entrainment of green sturgeon by turbines and other apparatus associated with hydrokinetic projects.

**Fisheries:** incidental capture of green sturgeon in fisheries. Incidental capture of adult green sturgeon occurs in several fisheries, including but not limited to the white sturgeon and salmonid recreational fisheries in the Sacramento and Feather rivers and Bay-Delta and bottom trawl fisheries (e.g., groundfish, California halibut) along the coast. Juvenile green sturgeon are incidentally captured in recreational fisheries in the lower Sacramento River and in shrimp trawl and commercial herring gillnet fisheries in the Bay-Delta. Although green sturgeon must be released, delayed migration of the adults and over-exertion by playing the fish could result in stress, injury, or immediate or delayed mortality from injury caused from the hooking or mishandling the fish. Some fish may be retained accidentally due to misidentification with white sturgeon and some may have increased predation risk by pinnipeds upon release. As of March 1, 2010, CDFW regulations prohibit fishing for any sturgeon species in the upper Sacramento River above Butte Bridge (Hwy 162).

**Global climate change:** large-scale changes to air and water temperature, precipitation, snow pack, and timing, frequency and magnitude of weather events. Climate change is expected to result in changes in precipitation from snow to rainfall, higher water temperatures, and increased frequency of high or low flow events. These changes could result in elevated river water temperatures that exceed the tolerable and/or optimal spawning and rearing temperatures for green sturgeon. Changing weather patterns, increased and fluctuating water temperature, and other changes may affect prey abundance, distribution, and community structure.

**Harvest of prey species:** legal and illegal harvest of prey species (e.g., crayfish and lamprey). Harvest may affect prey abundance and community structure. Additional prey species such as burrowing shrimp are actively eliminated to promote shellfish culture.

**Hatcheries:** artificial fish propagation facilities. The release of artificially propagated fish from hatcheries could introduce disease to wild green sturgeon populations.

Impoundments: impoundments created by dams (defined as rim, diversion, or low-head). In addition to dams delaying migration or blocking access to spawning habitat, impoundments can have several effects on in-river conditions. Impoundment outflows may be managed for power generation, flood control, water delivery, and water quality (e.g., saltwater intrusion). This can alter the natural hydrograph in the rivers, altering water temperatures and sediment composition and distribution. For example, impoundment outflows may be warmer (from the lake surface) or cooler (from the lake bottom) than unmanaged temperatures downstream of the dams. Sedimentdepleted conditions downstream of impoundments cause channel adjustment in the form of bank erosion, bed erosion, substrate coarsening, and channel planform change. Impoundments reduce or eliminate spring turbidity levels that may be beneficial to sturgeon spawning behavior, egg adhesion, egg development, and egg/larval camouflage from predators. Impoundments result in changes to channel morphology and pool structure by reducing sediment recruitment and transport magnitude and duration, minimizing flooding effects and natural redistribution of sediments and woody debris recruitment. Impoundments also affect conditions further downstream in bays, estuaries, and nearshore marine waters. For example, impoundments reduce sediment recruitment and transport magnitude and duration, minimize flooding effects and natural redistribution of sediments recruitment, and reduce turbidity by holding back sediment and acting as nutrient traps. This can result in changes to the bathymetry of bays and estuaries and nearshore habitats.

**In-water construction:** maintenance and construction of in-water structures including bridge piers and abutments, boat docks and marinas, diversions, and habitat restoration features. In-water construction activities may result in the release of fine to medium sized sediments and hazardous materials into the water column, such as fuels, oils, hydraulic fluids, chemicals, and contaminants in sediments (e.g., mercury or mercury byproducts). Sediment and contaminants released from in-water construction can affect egg survival and larval development, spawning and rearing habitat, water depth and bathymetry, and wetland and ecosystem function.

**In-water structures:** in-water structures include but are not limited to the Sacramento Deep Water Ship Channel locks and Delta Cross Channel gates, which may alter green sturgeon migration and/or be complete or partial migration barriers depending on operations. In-water structures in marine waters, like hydrokinetic projects and oil rigs, may alter prey availability by altering benthic communities.

**Marine mammals:** Steller and California sea lions and harbor seals. California sea lions have been observed preying on green sturgeon in the Columbia River and move upstream as far as Knights Landing (rkm 142) in the Sacramento River.

**Mitigation and restoration:** channel, floodplain, and tidal wetland habitat restoration projects. Restoration of channel and floodplain areas could alter geomorphology in green sturgeon spawning or holding habitat and restoration of tidal function in diked baylands could potentially modify the depth of foraging habitat in estuaries.

**Native species:** native fish (e.g., Sacramento suckers and pikeminnow), birds, and mammals. Native species can prey on green sturgeon eggs, larvae, and juveniles as well as compete with green sturgeon for rearing habitat and food.

**Non-native species:** non-native fish (e.g., striped and largemouth bass), crustaceans (e.g., red swamp crayfish and Siberian prawn), bivalves (e.g., Asian calm and overbite clam), and subaquatic vegetation (e.g., Brazilian waterweed and Japanese eelgrass). Non-native species can prey on green sturgeon eggs, larvae, and juveniles (e.g., striped bass); compete with green sturgeon for habitat and food; or compete with prey species and replace prey items of greater nutritional value. Some species may affect habitat quality, such as invasive subaquatic vegetation (e.g., *Egeria, Spartina alterniflora, Zostera japonica*) that alters wetland function and species composition in the Delta. Non-native species may also introduce disease to wild sturgeon populations, for example, through the introduction of non-native parasites or diseases, or the introduction of new host species.

**Oil and chemical spills:** oil and chemical spills associated with vessel accidents, railway transport of chemicals and subsequent derailment, or other activities that occur near water or in adjacent areas (e.g., vehicle fueling or heavy equipment operation). Exposure of green sturgeon to oil and chemicals could result in stress, injury, or death.

**Poaching:** the intentional and illegal harvest of subadult and adult green sturgeon for meat or roe. Poaching includes targeted harvest and intentional retention of incidentally caught fish.

**Point and non-point source contaminants:** point source contaminants include effluents from sewage treatment plants, timber mills, industrial facilities (e.g., power plants, LNG facilities, and desalination plants), and small fuel and chemical spills in and around water (e.g., riverbank docks, marinas, and other infrastructure). Non-point sources contaminants include runoff from urban areas (e.g., roads, building sites, domestic sewage, gardens, lawns, and parking lots), forests, agricultural areas, nurseries, landfills, livestock, and mining operations. Exposure of green sturgeon to contaminants could result in stress (reproductive or hormonal), injuries, and/or mortality. Contaminants also impact macroinvertebrate population and community structure (species presence/absence and overall numbers), affecting benthic prey resources for green sturgeon.

**Point and non-point source sediment:** point source sediments include but are not limited to fine-grain sediment from sewage treatment plants, timber mills, industrial discharges, riverbank docks, and marinas. Non-point source sediments include fine-grain sediment from urban and agricultural runoff, logging operations, burned areas, mining operations, and grazed lands. Input of fine grain material alters turbidity, substrate composition, and water depth/bathymetry. Removal of riparian vegetation also results in increased erosion and input of fine grain material into the water.

**Point and non-point source thermal effluence:** water inputs that increase ambient water temperatures. Point sources include Thermalito Afterbay Outlet, industrial and sewage treatment facility effluents, and power plants. Non-point sources include runoff from agricultural and urban areas (excluding bypasses).

**Point and non-point source turbidity:** temporary or ongoing sources of turbidity. Point source turbidity inputs include but are not limited to sewage treatment plants, timber mills, industrial discharges, riverbank docks, and marinas. Non-point turbidity sources include but are not limited

to runoff from urban and agricultural areas, forests, irrigated lands, landfills, livestock, mining operations, nurseries, orchards, and algal blooms.

**Sacramento River temperature management:** management of water temperature below Keswick Dam on the Sacramento River for species such as Chinook salmon. Reduced water temperature in the Sacramento River may influence green sturgeon spawning distribution (including spatial and temporal), egg incubation/development, and larval growth.

**Sand/gravel mining:** mining of aggregate material in channels of Suisun Bay, central San Francisco Bay, rivers and tributaries, adjacent riparian corridors, or off-site (e.g., at mine tailings along the Sacramento, Feather, and Yuba rivers). Mining activities may release fine-grain sediments into the water column and remove substrates that are used by green sturgeon for spawning. The removal of sand and gravel may also alter the prey base (similar to dredging activities), temporarily or permanently modify depth/bathymetry, and reduce sediment input to downstream areas.

**Scientific research activities:** scientific research activities including egg and larval sampling (e.g., egg mats, rotary screw traps, and D-nets), and capture and tagging of adults and juveniles (e.g., hook and line and gillnet). Collection, handling, and release of green sturgeon associated with research activities could cause stress, injuries, or mortality (i.e., exposure to pathogens during tagging, physiological stress due to handling, and suboptimal conditions during transport).

**Shoreline development:** activities include but are not limited to construction and short-term impacts of land clearing or excavation work and bank armoring to protect property adjacent to the river (e.g., waterfront homes, recreation areas, and agriculture). These activities result in erosion and release of sediments, loss or alteration of wetland function (e.g., by filling in the margins of bays for development or constructing/maintaining levees within the estuary/delta), or changes in sediment dynamics and turbidity (e.g., shoreline armoring).

**Upstream diversions:** diversions occurring upstream or outside of a given habitat unit. In the Sacramento River, the out of basin (i.e., Trinity River) transfer of water may increase flows. In other cases, diverting water upstream reduces the flow in the river and subsequently into estuaries and bays, altering channel hydraulics and sediment transport processes. Reduced flow may also result in increased water temperatures downstream of the diversion and ultimately influence water temperature in the nearshore marine environment.

Vessel propeller strikes: vessel propellers hitting green sturgeon and causing death or injury.

**Water quality:** water quality measures include but are not limited to temperature, turbidity, and dissolved oxygen. Poor water quality causes stress to green sturgeon, which may lead to reduced immune function and resilience and increased risk of disease or physiological stress. Poor water quality associated with low dissolved oxygen, contaminants, and temperature, may create migratory barriers (e.g., plumes of low dissolved oxygen, such as observed at the mouth of the Columbia River).

# Threat Analysis

Each recovery team member assessing a habitat unit received the same instructions, criteria, and relevant threat definition to conduct a detailed analysis of the current and potential future threats, following guidelines developed under Conservation Measures Partnership and Benetech's Miradi program (https://miradi.org/). Miradi utilizes categorical criteria (Very High, High, Medium, Low, or Not Applicable) for *Scope, Severity, Permanence,* and *Data Sufficiency,* and (Historical, Current, and Future) for *Threat Persistence* for each target (life stage) present as defined below. In addition to providing a score for each assessed parameter, each recovery team member was instructed to include a statement describing their rationale, including available reference citations that supported their decision. Each recovery team member conducted this initial assessment independently. The references considered by the recovery team members in conducting their assessment are included in this document and organized by habitat unit.

Following this initial assessment, the recovery team members assigned to each habitat unit met as a group to discuss their individual ratings and reach consensus on a final rating for each assessed parameter (*Scope, Severity, Permanence,* and *Data Sufficiency*). Once consensus was reached, the rankings for each assessed parameter was input into the "Miradi" software to derive an overall rating (Very High, High, Medium, or Low) for each specific threat for each life stage present in the habitat unit. Specifically, Miradi averaged the ratings for *Scope* and *Severity*, and that rating was then used with the *Permanence* rating to derive an overall rating for each specific threat. (Figure 1). *Data Sufficiency* scores were not used to derive a final ranking for each threat. Although Miradi also created a *summary rating* across all life stages for each specific threat, this was not used as we felt that it was important to consider the threat ratings for each life stage. Finally, the recovery team members met a final time to review the Miradi output and make any changes (which were few) to the final threats rankings.

# Definitions and Rating Criteria for Scope, Severity, Permanence, Data Sufficiency, and Threat Persistence

**Scope:** defined spatially as the proportion of the target that can reasonably be expected to be affected by the threat within fifty years given the continuation of current circumstances and trends. Two hierarchical methodologies can be used to reach this ranking. The primary approach rates the proportion of the population or life stage affected (% of fish). When these data are not known, a secondary approach can be used that rates the proportion of the habitat affected (% of habitat).

**Very High:** very widespread or pervasive in its scope, affecting the target across all or most (71-100%) of a designated ecosystem and life stage.

**High:** widespread in its scope, affecting the target across much (31-70%) of a designated ecosystem and life stage.

**Medium:** restricted in its scope, affecting the target across some (11-30%) of a designated ecosystem and life stage.

**Low:** very narrow in its scope, affecting the target across a small proportion (1-10%) of a designated ecosystem and life stage.

**Severity:** Within the scope, the level of damage to the target from the threat that can reasonably be expected given the continuation of current circumstances and trends. For green sturgeon, severity was estimated as the degree of reduction of life stage or habitat present within the designated region.

**Very High:** destroy or eliminate the target or reduce its population by 71-100% within fifty years.

High: seriously degrade habitat or reduce the population by 31-70% within fifty years.

**Medium:** moderately degrade habitat or reduce the population by 11-30% within fifty years.

Low: only slightly degrade habitat or reduce the population by 0-10% within fifty years.

**Permanence:** The degree to which the effects of a threat can be reversed and the target affected by the threat restored. Permanence refers to the effects of the threat on the target, not the threat itself. In other words, it is not a measure of how difficult it is to stop the threat, but rather to undo the stress caused by the threat on the target.

**Very High:** cannot be reversed and it is very unlikely the target can be restored, and/or it would take more than 100 years to achieve this (i.e., wetlands converted to a shopping center).

**High:** can technically be reversed and the target restored, but it is not practically affordable and/or it would take 21-100 years to achieve this (i.e., wetland converted to agriculture).

**Medium:** can be reversed and the target restored with a reasonable commitment of resources and/or within 6-20 years (i.e., ditching and draining of wetland).

**Low:** are easily reversible and the target can be easily restored at a relatively low cost and/or within 0-5 years (i.e., off-road vehicles trespassing in wetland).

Note: the permanence rating as specified incorporates both a temporal and irreversibility aspect with respect to prioritizing potential threats. For example, if a threat is imminent that will cause irreversible damage then it makes sense to prioritize that threat to avoid the impact. However, if the threat has already occurred and the irreversible damage has already taken place, then it may receive a lower priority.

**Data Sufficiency:** Data sufficiency refers to the quality of data available upon which to assign a ranking.

**High:** An abundance of data is available for the species and effects, and the scorer has no reservations in reaching a ranking decision.

**Medium:** Specific data are available for the species and effects, and a ranking decision can be assigned, but additional data are desired.

**Low:** Ranking decision is based on expert opinion, hypotheses, or suspicions based on biological concepts or inferences from data or information on other species or areas.

Threat Persistence: Is the threat historic, current, or future.

Historical: threats that occurred in the past and may or may not be occurring presently.

**Current:** threats occurring now.

Future: threats likely to affect green sturgeon over the next fifty years.

#### **Definitions for Overall Ratings**

The overall ratings for each threat were derived using the Miradi software, based on the consensus scores for Scope, Severity, and Permanence (Figure A-1). The overall ratings are defined below. Each overall rating level could result from several combinations of scores for the three parameters. A Very High or High score for scope/severity (together, called the "magnitude") or for permanence had a large influence on the overall rating.

- **Very High:** Threats with an overall rating of Very High had Very High scope/severity and Very High, High, or Medium permanence; or High scope/severity and Very High permanence.
- **High:** Threats with an overall rating of High had Very High scope/severity and Low permanence; High scope/severity and High or Medium permanence; or Medium scope/severity and Very High permanence.

#### Medium: Threats with an overall rating of Medium had High scope/severity and Low permanence; Medium scope/severity and High or Medium permanence; or Low scope/severity and Very High permanence.

#### Low: Threats with an overall rating of Low had Medium scope/severity and Low permanence; or Low scope/severity and High, Medium, or Low permanence.

(a)		Scope				
		Very High	High	Medium	Low	
Severity	Very High	Very High	High	Medium	Low	
	High	High	High	Medium	Low	
	Medium	Medium	Medium	Medium	Low	
	Low	Low	Low	Low	Low	

(b)		Permanence				
		Very High	High	Medium	Low	
Magnitude	Very High	Very High	Very High	Very High	High	
	High	Very High	High	High	Medium	
	Medium	High	Medium	Medium	Low	
	Low	Medium	Low	Low	Low	

Figure A-1. Miradi's rule-based system for deriving an overall rating for each threat based on Scope, Severity, and Permanence. (a) First, Miradi combines the ratings for scope and severity to get the overall threat magnitude rating for each threat. (b) Miradi then combines this overall threat magnitude rating from the first step with the permanence rating. Figure adapted from Appendix B in Foundations of Success. 2009. Conceptualizing and Planning Conservation Projects and Programs: A Training Manual. Foundations of Success, Bethesda, Maryland, USA (available at: http://www.fosonline.org/resource/conceptualizing-and-planning-manual).

# **References by Habitat Unit**

### Sacramento River Basin

Amoser, S. and F. Ladich. 2003. Diversity in Noise-Induced Temporary Hearing Loss in Otophysine Fishes. Journal of the Acoustical Society of America 113(4 Pt 1):2170-2179.

Andre, M., M. Sole, M. Lenoir, M. Durfort, C. Quero, A. Mas, A. Lombarte, M. van der Schaar, M. Lopez-Bejar, M. Morell, S. Zaugg, and L. Houegnigan. 2011. Low-Frequency Sounds Induce Acoustic Trauma in Cephalopods. Frontiers in Ecology and the Environment 9(9):489-493.

Apperson, K. A. and P. J. Anders. 1991. Kootenai River White Sturgeon Investigations and Experimental Culture. Report Prepared for the U.S. Department of Energy, Bonneville Power Administration, Division of Fish and Wildlife. Project No. 88-65. Contract No. DE-AI79-88BP93497, pp. 73.

Baker, D. W., B. P. Bledsoe, C. M. Albano, and N. L. Poff. 2011. Downstream Effects of Diversion Dams on Sediment and Hydraulic Conditions of Rocky Mountain Streams. River Research and Applications 27(3):388-401.

Bergen, K. 2010. Columbia Researchers Study Intersex Sturgeon in Missouri River. https://www.columbiamissourian.com/news/local/columbia-researchers-study-intersex-sturgeonin-missouri-river/article\_578002ab-ca93-5723-9299-8f82cf1d6fae.html. 5/4/2018.

Brown, K. 2007. Evidence of Spawning by Green Sturgeon, *Acipenser medirostris*, in the Upper Sacramento River, California. Environmental Biology of Fishes 79(3-4):297-303.

CALFED Science Program. 2018. Sound, Air and Light Barrier Keeps Chinook Out of Old River. CALFED Science Program Science News. http://science.calwater.ca.gov/publications/sci\_news\_1209\_bubble.html. 4/17/2018.

California Department of Fish and Game. 2004. Sacramento River Winter-Run Chinook Salmon. Biennial Report 2002 – 2003. Prepared for the Fish and Game Commission. California Department of Fish and Game, pp. 23.

California Department of Water Resources. 2004. Final Report - Characterization of Cold Water Pool Availability in the Thermalito Afterbay Oroville Facilities P-2100 Relicensing. Oroville Facilities Relicensing Team, SP-F3.1, Task 4b. Oroville Facilities Relicensing. FERC Project No. 2100.

California Department of Water Resources. 2004. SPW6. Project Effects on Temperature Regime Oroville Facilities Relicensing FERC Project No. 2100. California Department of Water Resources, pp. 151. California Department of Water Resources. 2008. Sacramento Stormwater Quality Partnership Annual Report 2007-2008. October 2008.

California Sportfishing Protection Alliance. 2018. California Sportfishing Protection Alliance. http://calsport.org/about/whois.php. 5/8/2018.

Caroffino, D. C., T. M. Sutton, R. F. Elliott, and M. C. Donofrio. 2010. Predation on Early Life Stages of Lake Sturgeon in the Peshtigo River, Wisconsin. Transactions of the American Fisheries Society 139(6):1846-1856.

Carrasquero, J. 2001. Executive Summary: Overwater Structures: Freshwater Issues. pp. 129.

Davidson, J., J. Bebak, and P. Mazik. 2009. The Effects of Aquaculture Production Noise on the Growth, Condition Factor, Feed Conversion, and Survival of Rainbow Trout, *Oncorhynchus mykiss*. Aquaculture 288(3-4):337-343.

Dredging Today. 2011. USA Mercury Contamination in California's South Yuba River Causes Concern. https://www.dredgingtoday.com/2011/01/26/usa-mercury-contamination-in-california%E2%80%99s-south-yuba-river-causes-concern. 5/7/2018.

Drennan, J. D., S. E. LaPatra, J. T. Siple, S. Ireland, and K. D. Cain. 2006. Transmission of White Sturgeon Iridovirus in Kootenai River White Sturgeon *Acipenser transmontanus*. Diseases of Aquatic Organisms 70:37-45.

DuBois, J., T. Matt, and T. MacColl. 2011. 2010 Sturgeon Fishing Report Card: Preliminary Data Report. California Department of Fish and Game, pp. 14.

Engas, A., S. Lokkeborg, E. Ona, and A. V. Soldal. 1996. Effects of Seismic Shooting on Local Abundance and Catch Rates of Cod (*Gadus morhua*) and Haddock (*Melanogrammus aeglefinus*). Canadian Journal of Fisheries and Aquatic Sciences 53(10):2238-2249.

Enger, P. S. 1981. Frequency Discrimination in Teleosts -- Central or Peripheral? Pages 243-255 *in* Hearing and Sound Communication in Fishes. Proceedings in Life Sciences. Springer, New York.

Environment and Climate Change Canada. 2018. The Effects of Non-Point Source Pollution in Small Urban and Agricultural Streams Interpretive Report. Environment and Climate Change. http://www.ec.gc.ca/pabg-gbap/Default.asp?lang=En&n=730672A0-1. 4/17/2018.

Feyrer, F., T. Sommer, and W. Harrell. 2006. Importance of Flood Dynamics Versus Intrinsic Physical Habitat in Structuring Fish Communities: Evidence from Two Adjacent Engineered Floodplains on the Sacramento River, California. North American Journal of Fisheries Management 26(2):408-417. Gadomski, D. M. and M. J. Parsley. 2005. Effects of Turbidity, Light Level, and Cover on Predation of White Sturgeon Larvae by Prickly Sculpins. Transactions of the American Fisheries Society 134(2):369-374.

Gale, R., D. Papoulias, and D. Tillitt. 2018. USGS Columbia Environmental Research Center Project: Reproductive Effects of Contaminants on Shovelnose Sturgeon. http://www.cerc.usgs.gov/Projects.aspx?ProjectId=141. 4/17/2018.

Gisbert, E. and P. Williot. 1997. Larval Behaviour and Effect of the Timing of Initial Feeding on Growth and Survival of Siberian Sturgeon (*Acipenser baeri*) Larvae under Small Scale Hatchery Production. Aquaculture 156:63-76.

Gray, L. 2004. Changes in Water Quality and Macroinvertebrate Communities Resulting from Urban Stormflows in the Provo River, Utah, U.S.A. Hydrobiologia 518(1-3):33-46.

Hastings, M. C. and A. N. Popper. 2005. Effects of Sound on Fish. California Department of Transportation, pp. 82.

Hastings, M. C., A. N. Popper, J. J. Finneran, and P. J. Lanford. 1996. Effects of Low-Frequency Underwater Sound on Hair Cells of the Inner Ear and Lateral Line of the Teleost Fish *Astronotus ocellatus*. Journal of the Acoustical Society of America 3:1759-1766.

Hedrick, R. and S. E. LaPatra. 2001. Characteristics of Viral Diseases of Sturgeon and Their Impacts on Aquaculture, Conversation, and International Trade.

Herren, J. R. and S. Kawasaki. 2001. Inventory of Water Diversions in Four Geographic Areas in California's Central Valley. Fish Bulletin 2:343-355.

Hu, J., Z. Zhang, Q. Wei, H. Zhen, Y. Zhao, H. Peng, Y. Wan, J. P. Giesy, L. Li, and B. Zhang. 2009. Malformations of the Endangered Chinese Sturgeon, *Acipenser sinensis*, and Its Causal Agent. Proceedings of the National Academy of Sciences of the United States 106(23):9339-9344.

IEA Hydropower. IEA Hydropower Implementing Agreement Annex VIII - Hydropower Good Practices: Environmental Mitigation Measures and Benefits Case Study 05-08. U. S. Bureau of Reclamation, pp. 12.

IPCC (Intergovernmental Panel on Climate Change). 2007. Climate Change 2007 the Physical Science Basis, Contribution of Working Group I to the Fourth Assessment. Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge.

Jackson, J. R., A. J. Van De Valk, T. E. Brooking, O. A. Van Keeken, and L. G. Rudstam. 2002. Growth and Feeding Dynamics of Lake Sturgeon, *Acipenser fulvescens*, in Oneida Lake, New York: Results from the First Five Years of a Restoration Program. Journal of Applied Ichthyology 18(4-6):439-443. Jeffries, K. M., L. J. Jackson, M. G. Ikonomou, and H. R. Habibi. 2010. Presence of Natural and Anthropogenic Organic Contaminants and Potential Fish Health Impacts Along Two River Gradients in Alberta, Canada. Environmental Toxicology and Chemistry 29(10):2379-2387.

Kaufman, R. C., A. G. Houck, and J. J. Cech. 2008. Effects of Dietary Selenium and Methylmercury on Green and White Sturgeon Bioenergetics in Response to Changed Environmental Conditions *in* 5th Biennial CALFED Science Conference.

Knight, S. S. and K. L. Boyer. 2007. Effects of Conservation Practices on Aquatic Habitats and Fauna. pp. 19.

Knudsen, F. R., P. S. Enger, and O. Sand. 1994. Avoidance Responses to Low-Frequency Sound in Downstream Migrating Atlantic Salmon Smolt, *Salmo salar*. Journal of Fish Biology 45(2):227-233.

Kynard, B., E. Parker, and T. Parker. 2005. Behavior of Early Life Intervals of Klamath River Green Sturgeon, *Acipenser medirostris*, with a Note on Body Color. Environmental Biology of Fishes 72(1):85-97.

Lawless, A. S. 2008. Effects of Shoreline Development and Oyster Reefs on Benthic Communities in Lynnhaven, Virginia. Master's Thesis. The College of William and Mary in Virginia.

Lessard, J. L. and D. B. Hayes. 2003. Effects of Elevated Water Temperature on Fish and Macroinvertebrate Communities Below Small Dams. River Research and Applications 19(7):721-732.

Lisle, T. E. 1982. Effects of Aggradation and Degradation on Riffle-Pool Morphology in Natural Gravel Channels, Northwestern California. Water Resources Research 18(6):1643-1651.

Loew, E. R. and A. J. Sillman. 1998. An Action Spectrum for the Light-Dependent Inhibition of Swimming Behavior in Newly Hatched White Sturgeon, *Acipenser transmontanus*. Vision Research 38(1):111-114.

Lower Yuba River Accord River Management Team Planning Group. 2010. Lower Yuba River Water Temperature Objectives Technical Memorandum. pp. 75.

Mayfield, R. B. and J. J. Cech. 2004. Temperature Effects on Green Sturgeon Bioenergetics. Transactions of the American Fisheries Society 133(4):961-970.

McAdam, S. O. 2011. Effects of Substrate Condition on Habitat Use and Survival by White Sturgeon (*Acipenser transmontanus*) Larvae and Potential Implications for Recruitment. Canadian Journal of Fisheries and Aquatic Sciences 68(5):812-822.

McCauley, R. D., J. Fewtrell, and A. N. Popper. 2003. High Intensity Anthropogenic Sound Damages Fish Ears. Journal of the Acoustical Society of America 113(1):638-642.

Meier, W., C. Bonjour, A. Wuest, and P. Reichert. 2003. Modeling the Effect of Water Diversion on the Temperature of Mountain Streams. Journal of Environmental Engineering 129(8):755-764.

Merz, J. E. and L. K. O. Chan. 2005. Effects of Gravel Augmentation on Macroinvertebrate Assemblages in a Regulated California River. River Research and Applications 21(1):61-74.

Mora, E. A., S. T. Lindley, D. L. Erickson, and A. P. Klimley. 2009. Do Impassable Dams and Flow Regulation Constrain the Distribution of Green Sturgeon in the Sacramento River, California? Journal of Applied Ichthyology 25:39-47.

Moyle, P. B. 2002. Inland Fishes of California. University of California Press, Berkeley and Los Angeles.

National Marine Fisheries Service. 2009a. NMFS Biological and Conference Opinion on the Long-Term Operations of the Central Valley Project and State Water Project. pp. 844.

Nguyen, R. M. and C. E. Crocker. 2007. The Effects of Substrate Composition on Foraging Behavior and Growth Rate of Larval Green Sturgeon, *Acipenser medirostris*. Environmental Biology of Fishes 76(2-4):129-138.

Ocean Mammalist Institute. 2018. Intense Ocean Noise Effects on Fish www.oceanmammalinst.com/ocean-noise-fish-fact-sheet.htm. 4/17/2018.

Ostrand, K. G., W. G. Simpson, C. D. Suski, and A. J. Bryson. 2009. Behavioral and Physiological Response of White Sturgeon to an Electrical Sea Lion Barrier System. Marine and Coastal Fisheries 1(1):363-377.

Pasternack, G. 2009. Current Status of an on-Going Gravel Injection Experiment on the Lower Yuba River, CA. pp. 26.

Pearson, W. H., J. R. Skalski, and C. I. Malme. 1992. Effects of Sound from a Geophysical Survey Device on Behavior of Captive Rockfish. Canadian Journal of Fisheries and Aquatic Science 49:1343-1356.

Poddubny, A. G. 1967. Sonic Tags and Floats as a Means of Studying Fish Response to Natural Environmental Changes to Fishing Gears. Pages 793-802 *in* Conference on Fish Behaviour in Relation to Fishing Techniques and Tactics, Bergen, Norway FAO, Rome.

Popper, A. N. 2003. Effects of Anthropogenic Sounds on Fishes. Fisheries Research 28(10):24-31.

Popper, A. N. and N. L. Clarke. 1976. The Auditory System of the Goldfish (*Carassius auratus*): Effects of Intense Acoustic Stimulation. Comparative Biochemistry and Physiology 53(1):11-18.

Popper, A. N. and R. R. Fay. 1973. Sound Detection and Processing by Teleost Fishes: A Critical Review. Journal of the Acoustical Society of America 53(6):1515-1529.

Popper, A. N., R. R. Fay, C. Platt, and O. Sand. 2003. Sound Detection Mechanisms and Capabilities of Teleost Fishes. Pages 3-38 *in* Sensory Processing in Aquatic Environments, S. P. Collin and N. J. Marshall, editors. Springer.

Poytress, W. R., J. J. Gruber, D. A. Trachtenbarg, and J. van Eenennaam. 2009. 2008 Upper Sacramento River Green Sturgeon Spawning Habitat and Larval Migration Surveys. U.S. Fish and Wildlife Service and University of California Davis, pp. 1-52.

Poytress, W. R., J. J. Gruber, and J. P. Van Eenennaam. 2010. 2009 Upper Sacramento River Green Sturgeon Spawning Habitat and Larval Migration Surveys. U.S. Fish and Wildlife Service and University of California Davis, pp. 1-48.

Poytress, W. R., J. J. Gruber, and J. P. Van Eenennaam. 2011. 2010 Upper Sacramento River Green Sturgeon Spawning Habitat and Larval Migration Surveys U.S. Fish and Wildlife Service and University of California Davis, pp. 1-48.

Pratt, J. M., R. A. Coler, and P. J. Godfrey. 1981. Ecological Effects of Urban Stormwater Runoff on Benthic Macroinvertebrates Inhabiting the Green River, Massachusetts. Hydrobiologia 83(1):29-42.

Riverkeeper Inc. 2010. Scoping Comments for the Champlain Hudson Power Express Transmission Line Project Draft Environmental Impact Statement, Docket No. PP-362.

Sacramento River Advisory Council. 2003. Sacramento River Conservation Area Forum Handbook. The Resources Agency, pp. 239.

Sacramento River Watershed Program. 2018. Sacramento Valley Subregion Sacramento River Basin Watersheds. http://www.sacriver.org/aboutwatershed/roadmap/watersheds/sacramento-valley-subregion. 4/17/2018.

Sand, O., P. S. Enger, H. E. Karlsen, F. R. Knudsen, and T. Kvernstuen. 2000. Avoidance Responses to Infrasound in Downstream Migrating European Silver Eels, *Anguilla*. Environmental Biology of Fishes 57:327-336.

Schulz, R., G. Thiere, and J. M. Dabrowski. 2002. A Combined Microcosm and Field Approach to Evaluate the Aquatic Toxicology of Azinphosmethyl to Stream Communities. Environmental Toxicology and Chemistry 21(10):2172-2178.

Seelye, J. G. and M. J. Mac. 1984. Bioaccumulation of Toxic Substances Associated with Dredging and Dredged Material Disposal. Chicago, Illinois.

Silvestre, F., J. Linares-Casenave, S. I. Doroshov, and D. Kultz. 2010. A Proteomic Analysis of Green and White Sturgeon Larvae Exposed to Heat Stress and Selenium. Science of the Total Environment 408(16):3176-3188.

Smith, M. E., A. S. Kane, and A. N. Popper. 2004. Noise-Induced Stress Response and Hearing Loss in Goldfish (*Carassius auratus*). The Journal of Experimental Biology 207(Pt 3):427-435.

Sommer, T., W. Harrell, M. Nobriga, R. Brown, P. Moyle, W. Kimmerer, and L. Schemel. 2011. California's Yolo Bypass: Evidence That Flood Control Can Be Compatible with Fisheries, Wetlands, Wildlife, and Agriculture. Fisheries 26(8):6-16.

Stillwater Sciences. 2006. Sacramento River Ecological Flows Study: State of the System Report. Prepared for The Nature Conservancy.

Sun, Y., Y. Song, J. Zhao, J. Chen, Y. Yuan, S. Jiang, and D. Zhang. 2001. Effect of Drilling Noise and Vibration on Growth of Carp (*Cyprinus carpio*) by Cut-Fin Marking. Marine Fish Research/Haiyang Shiuchan Yanjiu 22(1):62-68.

Sverdrup, A., E. Kjellsby, P. G. Kruger, R. Floysand, F. R. Knudsen, P. S. Enger, G. Serckhanssen, and K. B. Helle. 1994. Effects of Experimental Seismic Shock on Vasoactivity of Arteries, Integrity of the Vascular Endothelium and on Primary Stress Hormones of the Atlantic Salmon. Journal of Fish Biology 45(6):973-995.

Systech Water Resources, I. 2001. Task 3 Technical Memorandum Analytical Modeling of the Sacramento River. I. Systech Water Resources, pp. 135.

Thompson, A. M., K. Kim, and A. J. Vandermuss. 2008. Thermal Characteristics of Stormwater Runoff from Asphalt and Sod Surfaces. Journal of the American Water Resources Association 44(5):1325-1336.

Tolimieri, N., A. Jeffs, and J. C. Montgomery. 2002. Ambient Sound as a Cue for Navigation by the Pelagic Larvae of Reef Fishes. Bioacoustics 207:219-224.

Tucker, M. E., C. M. Williams, and R. R. Johnson. 1998. Abundance, Food Habits and Life History Aspects of Sacramento Squawfish and Striped Bass at the Red Bluff Diversion Complex, Including the Research Pumping Plant, Sacramento River, California, 1994-1996. U.S. Fish and Wildlife Service, pp. 63.

U.S. Army Corps of Engineers. 2018. Portland District Missions Environmental Stewardship. http://www.nwp.usace.army.mil/environment/docs/afep. 5/8/2018.

U.S. Bureau of Reclamation. 2004. Biological Assessment on the Continued Long-Term Operations of the Central Valley Project and the State Water Project. U.S. Bureau of Reclamation, pp. 108.

U.S. Bureau of Reclamation. 2010. Sacramento River Salmonid Spawning Gravel Augmentation Program Gets Under Way.

http://www.sustainablecitynetwork.com/topic\_channels/environmental/article\_ddf575ca-aeed-11df-980c-0017a4a78c22.html. 4/17/2018.

U.S. Bureau of Reclamation. 2011. Secure Water Act Section 9503(C) – Reclamation Climate Change and Water, Report to Congress. U.S. Department of the Interior, pp. 226.

U.S. Environmental Protection Agency. 2006. Iron Mountain Mine Case Study, Success through Planning, Partnerships, and Perseverance. U.S. Environmental Protection Agency, pp. 17.

U.S. Environmental Protection Agency. 2017. Caddis Volume 2 Sources, Stressors and Responses: Temperature. https://www.epa.gov/caddis-vol2/caddis-volume-2-sources-stressors-responses-temperature. 5/7/2018.

U.S. Fish and Wildlife Service. 2004 Impacts of Riprapping to Aquatic Organisms and River Functioning, Lower Sacramento River, California. U.S. Fish and Wildlife Service, pp. 44.

U.S. Fish and Wildlife Service, U.S. Bureau of Reclamation, and H. V. T. T. County. 2018. Trinity River Fishery Restoration, Supplemental Environmental Impact Statement/Environmental Report, Appendix G-3, OCAP BA Project Description. U.S. Bureau of Reclamation, U.S. Fish and Wildlife Service, pp. 108.

Upper Columbia White Sturgeon Recovery Initiative. 2018. Documents. http://www.nwp.usace.army.mil/environment/docs/afep/. 5/8/2018.

USCODE. 2010. Title 33 Navigation and Navigable Waters. Section 1329. Nonpoint Source Management Programs. pp. 457-462.

Van Eenennaam, J. P., J. Linares-Casenave, X. Deng, and S. I. Doroshov. 2005. Effect of Incubation Temperature on Green Sturgeon Embryos, *Acipenser medirostris*. Environmental Biology of Fishes 72:145-154.

Vogel, D. 1995. Losses of Young Anadromous Salmonids at Water Diversions on the Sacramento and Mokelumne Rivers.

Vogel, D. 2008. Evaluation of Adult Sturgeon Migration at the Glenn-Colusa Irrigation District Gradient Facility on the Sacramento River. Natural Resource Scientists, Inc.

Washington Department of Ecology. Wetlands: Nature's Water Filters, Sponges and Nurseries. https://ecology.wa.gov/Water-Shorelines/Wetlands/Wetlands-overview. 5/7/2018.

Wohl, E. and S. Rathburn. 2003. Mitigation of Sedimentation Hazards Downstream from Reservoirs. International Journal of Sediment Research 18:97-106.

Wysocki, L. E., J. P. Dittami, and F. Ladich. 2006. Ship Noise and Cortisol Secretion in European Freshwater Fishes. Biological Conservation 128(4):501-508.

## San Francisco Bay Delta and Estuary

Bakke, A. M., D. H. Tashjian, C. F. Wang, S. H. Lee, S. C. Bai, and S. S. O. Hung. 2010. Competition between Selenomethionine and Methionine Absorption in the Intestinal Tract of Green Sturgeon (*Acipenser medirostris*). Aquatic Toxicology 96:62-69.

Boehlert, G. W. and A. B. Gill. 2010. Environmental and Ecological Effects of Ocean Renewable Energy Development Current Synthesis. Oceanography 23(2):68-81.

Boehlert, G. W., G. R. McMurray, and C. E. Tortorici. 2008. Ecological Effects of Wave Energy Development in the Pacific Northwest, A Scientific Workshop. National Marine Fisheries Service, NOAA Technical Memorandum NMFS-F/SPO-92, pp. 186.

Gadomski, D. M. and M. J. Parsley. 2005. Effects of Turbidity, Light Level, and Cover on Predation of White Sturgeon Larvae by Prickly Sculpins. Transactions of the American Fisheries Society 134(2):369-374.

Heppell, S. S. 2007. Elasticity Analysis of Green Sturgeon Life History. Environmental Biology of Fishes 79(3-4):357-368.

Kelly, J. T., A. P. Klimley, and C. E. Crocker. 2007. Movements of Green Sturgeon, *Acipenser medirostris*, in the San Francisco Bay Estuary, California. Environmental Biology of Fishes 79(3-4):1-44.

Kenny, A. J. and H. L. Rees. 1994. The Effects of Marine Gravel Extraction on the Macrobenthos - Early Post-Dredging Recolonization. Marine Pollution Bulletin 28(7):442-447.

McKechnie, R. J. and R. B. Fenner. 1971. Food Habits of White Sturgeon (*Acipenser transmontanus*) in San Pablo and Suisun Bays, California. California Department of Fish and Game, pp. 209-212.

Mora, E. A., S. T. Lindley, D. L. Erickson, and A. P. Klimley. 2009. Do Impassable Dams and Flow Regulation Constrain the Distribution of Green Sturgeon in the Sacramento River, California? Journal of Applied Ichthyology 25:39-47.

Muir, W. D., R. L. Emmett, and R. J. McConnell. 1986. Diet of Juvenile and Subadult White Sturgeon in the Lower Columbia River and Its Estuary. National Marine Fisheries Service, pp. 19.

National Research Council. 2011. A Review of the Use of Science and Adaptive Management in California's Draft Bay Delta Conservation Plan.

Polagye, B., M. Kawase, and P. Malte. 2009. In-Stream Tidal Energy Potential of Puget Sound,

Washington. Proceedings of the Institution of Mechanical Engineers, Part A: Journal of Power and Energy 223(5):571-587.

Radtke, L. D. 1966. Distribution of Smelt, Juvenile Sturgeon, and Starry Flounder in the Sacramento-San Joaquin Delta with Observations on Food of Sturgeon. California Department of Fish and Game, pp. 115-119.

Schreiber, M. R. 1962. Observations on the Food Habits of Juvenile White Sturgeon. California Department of Fish and Game 48:79-80.

Schultz, I. R., W. J. Pratt, D. L. Woodruff, G. Roesijadi, and K. E. Marshall. 2010. Effects of Electromagnetic Fields on Fish and Invertebrates Task 2.1.3: Effects on Aquatic Organisms – Fiscal Year 2010 Progress Report Environmental Effects of Marine and Hydrokinetic Energy.

Silvestre, F., J. Linares-Casenave, S. I. Doroshov, and D. Kultz. 2010. A Proteomic Analysis of Green and White Sturgeon Larvae Exposed to Heat Stress and Selenium. Science of the Total Environment 408(16):3176-3188.

U.S. Fish and Wildlife Service. 2008. Species at Risk from Selenium Exposure in the San Francisco Estuary. U.S. Department of the Interior, pp. 85.

Van Dalfsen, J. A. and K. Essink. 2001. Benthic Community Response to Sand Dredging and Shoreface Nourishment in Dutch Coastal Waters. Senckenbergiana maritima 31(2):329-332.

### Coastal Bays and Estuaries

Cohen, A. 2005. Sturgeon Poaching and Black Market Caviar: A Case Study. Underwater Naturalist 27(1):23-26.

Moser, M. L., K. Patten, S. C. Corbett, B. E. Feist, and S. T. Lindley. 2017. Abundance and Distribution of Sturgeon Feeding Pits in a Washington Estuary. Environmental Biology of Fishes 100(5):597-609.

Parsley, M. J., N. D. Popoff, and J. G. Romine. 2011. Short-Term Response of Subadult White Sturgeon to Hopper Dredge Disposal Operations. North American Journal of Fisheries Management 31(1):1-11.

Thrush, S. F. and P. K. Dayton. 2002. Disturbance to Marine Benthic Habitats by Trawling and Dredging: Implications for Marine Biodiversity. Annual Review of Ecology and Systematics 33(1):449-473.

## Nearshore Marine

Adams, P. B., C. Grimes, J. E. Hightower, S. T. Lindley, M. L. Moser, and M. J. Parsley. 2007. Population Status of North American Green Sturgeon, *Acipenser medirostris*. Environmental Biology of Fishes 79(3-4):339-356.

Ayaz, A., V. Ünal, D. Acarli, and U. Altinagac. 2010. Fishing Gear Losses in the Gökova Special Environmental Protection Area (SEPA), Eastern Mediterranean, Turkey. Journal of Applied Ichthyology 26(3):416-419.

Bellman, M. A., E. Heery, and J. Majewski. 2010. Observed and Estimated Total Bycatch of Green Sturgeon in the 2002-2009 U.S. West Coast Fisheries. National Marine Fisheries Service, pp. 38.

Boysen, K. A. and J. J. Hoover. 2009. Swimming Performance of Juvenile White Sturgeon *(Acipenser transmontanus)*: Training and the Probability of Entrainment Due to Dredging. Journal of Applied Ichthyology 25:54-59.

Brown, J. J. and G. W. Murphy. 2010. Atlantic Sturgeon Vessel-Strike Mortalities in the Delaware Estuary. Fisheries 35(2):72-83.

Buell, J. W. 1992. Fish Entrainment Monitoring of the Western-Pacific Dredge R W Lofgren During Operations Outside the Preferred Work Period. Buell & Associates, Inc.

Cohen, A. 2005. Sturgeon Poaching and Black Market Caviar: A Case Study. Underwater Naturalist 27(1):23-26.

Dernie, K. M., M. J. Kaiser, and R. M. Warwick. 2003. Recovery Rates of Benthic Communities Following Physical Disturbance. Journal of Animal Ecology 72:1043-1056.

Erickson, D. L. and J. E. Hightower. 2007. Oceanic Distribution and Behavior of Green Sturgeon. American Fisheries Society Symposium 56:197-211.

Gilardi, K. V., D. Carlson-Bremer, J. A. June, K. Antonelis, G. Broadhurst, and T. Cowan. 2010. Marine Species Mortality in Derelict Fishing Nets in Puget Sound, WA and the Cost/Benefits of Derelict Net Removal. Marine Pollution Bulletin 60(3):376-382.

Good, T. P., J. A. June, M. A. Etnier, and G. Broadhurst. 2010. Derelict Fishing Nets in Puget Sound and the Northwest Straits: Patterns and Threats to Marine Fauna. Marine Pollution Bulletin 60(1):39-50.

Hannah, R. W., S. J. Parker, and T. V. Buell. 2011. Evaluation of a Selective Flatfish Trawl and Diel Variation in Rockfish Catchability as Bycatch Reduction Tools in the Deepwater Complex Fishery Off the U.S. West Coast. North American Journal of Fisheries Management 25(2):581-593.

Hatin, D., S. Lachance, and D. Fournier. 2007. Effect of Dredged Sediment Deposition on Use by Atlantic Sturgeon and Lake Sturgeon at an Open-Water Disposal Site in the St. Lawrence Estuarine Transition Zone. American Fisheries Society Symposium 56:235-255.

Heery, E., M. A. Bellman, and J. Hastie. 2009. Observed and Estimated Total Bycatch of Green Sturgeon in the 2002-2007 U.S. West Coast Groundfish Fisheries. Northwest Fisheries Science Center (NWFSC), pp. 25.

Keller, A. A., B. H. Horness, E. L. Fruh, V. H. Simon, V. J. Tuttle, K. L. Bosley, J. C. Buchannan, D. J. Kamikawa, and J. R. Wallace. 2008. The 2005 U.S. West Coast Bottom Trawl Survey of Groundfish Resources Off Washington, Oregon, and California: Estimates of Distribution, Abundance, and Length Composition. National Marine Fisheries Service, NOAA Technical Memorandum NMFS-NWFSC-93, pp. 1-153.

Lovell, J. M., M. M. Findlay, R. M. Moate, J. R. Nedwell, and M. A. Pegg. 2005. The Inner Ear Morphology and Hearing Abilities of the Paddlefish (*Polyodon spathula*) and the Lake Sturgeon (*Acipenser fulvescens*). Science Direct 142(3):286-296.

Mitchell, G. H., C. J. Lazauski, N. Lange, C. P. Damon, and J. M. Frederickson. 2008. Assessing Risk from Underwater Sound Using Simulation of Live-Fire/Antisubmarine Warfare Naval Exercises in the Presence of Protected Marine Species. Bioacoustics 17(1-3):257-259.

National Marine Fisheries Service. 2011. Federal Pacific Coast Groundfish Regulations for Commercial and Recreational Fishing 3-200 Nautical Miles Off Washington, Oregon, and California. Northwest Region pp. 343.

National Marine Fisheries Service and Northwest Fisheries Science Center. 2008. Data Report and Summary Analyses of the West Coast California Halibut Trawl Fishery. Seattle, Washington.

National Marine Fisheries Service and Northwest Fisheries Science Center (NWFSC). 2010. Data Report and Summary Analyses of the U.S. West Coast California Halibut Trawl Fishery. U.S. Department of Commerce.

Pacific Fisheries Information Network (PacFIN). 2000-2011. Gillnet/trammel net data to look at green sturgeon interactions. Pacific States Marine Fisheries Commission, Portland, Oregon.

Pacific Fishery Management Council (PFMC). 2011. Active West Coast Hydrokinetic Projects. Supplemental Information Report 4. pp. 5.

Parsley, M. J., N. D. Popoff, and J. G. Romine. 2011. Short-Term Response of Subadult White Sturgeon to Hopper Dredge Disposal Operations. North American Journal of Fisheries Management 31(1):1-11.

Pikitch, E. K., D. L. Erickson, and J. R. Wallace. 1988. An Evaluation of the Effectiveness of Trip Limits as a Management Tool. NWAFC Processed Report 88-27. National Marine Fisheries Service Northwest and Alaska Fisheries Center.

Reedsport OPT Wave Park. 2008a. Reedsport OPT Wave Park, FERC Project No. 12713. Draft Application for a Major License. An Oregon Corporation, pp. 55. Reedsport OPT Wave Park (ROPT). 2008b. Reedsport OPT Wave Park, FERC No. 12713, Aquatic Species Sub Group. Draft Issue Assessment Issue No. 2 EMF. Rev 1 - January 11, 2008. pp. 19.

Suuronen, P. and D. L. Erickson. 2010. Mortality of Animals That Escape Fishing Gears or Are Discarded after Capture: Approaches to Minimize Mortality. Page 54 *in* Behavior of Marine Fishes: Capture Processes and Conservation Challenges, P. He, editor. Wiley-Blackwell.

Thrush, S. F. and P. K. Dayton. 2002. Disturbance to Marine Benthic Habitats by Trawling and Dredging: Implications for Marine Biodiversity. Annual Review of Ecology and Systematics 33(1):449-473.

Tyack, P. L. 2008. Implications for Marine Mammals of Large-Scale Changes in the Marine Acoustic Environment. Journal of Mammalogy 89(3):549-558.

Vanderlaan, A. S. M., J. J. Corbett, S. L. Green, J. A. Callahan, C. Wang, R. D. Kenney, C. T. Taggart, and J. Firestone. 2009. Probability and Mitigation of Vessel Encounters with North Atlantic Right Whales. Endangered Species Research 6:273-285.

Woodbury, D. P. and J. H. Stadler. 2008. A Proposed Method to Assess Physical Injury to Fishes from Underwater Sound Produced During Pile Driving. Bioacoustics 17(1-3):289-291.